

The role of selected anions in the photodegradation of vancomycin in the aquatic environment

Paulina Sowik^{1*} , Katarzyna Kowalska^{1,2}, Ewa Felis^{1,2}

¹ Silesian University of Technology, Faculty of Energy and Environmental Engineering, Environmental Biotechnology Department, Akademicka 2, 44-100, Gliwice, Poland

² Silesian University of Technology, Biotechnology Centre, B. Krzywoustego 8, 44-100, Gliwice, Poland

* Corresponding author's e-mail: psowik@polsl.pl

ABSTRACT

The efficiency of micropollutant photodegradation in aquatic matrices depends on several factors, including the presence of anions. They can act as radical scavengers and therefore inhibit the removal rate of micropollutants. Understanding the mechanisms of these interactions is crucial for the further development of advanced treatment technologies of water. Taking this into account, the main objective of the present study was to evaluate the effect of selected anions, such as Cl^- , SO_4^{2-} , NO_3^- , and HCO_3^- , on the photolytic and photocatalytic degradation of vancomycin (VAN) in an aqueous matrix. The experiments were carried out in a solar simulator (Solarbox) using TiO_2 and ZnO as photocatalysts and Milli-Q water as an aquatic matrix. Additional tests were performed in the presence of peroxymonosulfate (PMS) and sulfate ions (SO_4^{2-}) to assess their influence on VAN degradation efficiency in the sulfate radical-based advanced oxidation process. The results demonstrated that vancomycin photodegradation efficiency was significantly influenced by the type of anions present in the solution. The scavenging effects of reactive oxygen species increase in the following order: $\text{HCO}_3^- > \text{NO}_3^- > \text{Cl}^- > \text{SO}_4^{2-}$. Moreover, the presence of sulfate ions adversely affected the PMS-assisted photodegradation, with a marked decrease in efficiency in most cases, except for TiO_2 -based photocatalysis, where comparable removal was still achieved.

Keywords: photodegradation, vancomycin, anions, peroxymonosulfate.

INTRODUCTION

The presence of antibiotics in the environment might have adverse effects on living organisms and lead to their bioaccumulation in soil or in the organisms' tissues (Felis et al., 2019; Zhang et al., 2022). However, the most hazardous effect of the presence of antibiotics in different environmental matrices is contributing to the emergence and spread of the phenomenon of antimicrobial resistance (AMR). European Centre for Disease Prevention and Control (ECDC) estimates that each year throughout countries of European Union (EU), Iceland and Norway, more than 35 000 deaths are related to infections caused by antibiotic-resistant bacteria (ARB) (ECDC, 2024). For decades, vancomycin (VAN) has been classified as a “last-resort antibiotic” (i.e.

“reserve”), although it is currently classified as a “watch” according to the WHO classification system. It should be mentioned here that the group of “reserve antibiotics” includes medicines that should be reserved for treatment of confirmed or suspected infections due to multidrug-resistant organisms and only in hospital conditions (WHO, 2023). The “Watch list” comprises antibiotic classes with a high potential to drive antimicrobial resistance. It includes many of the highest-priority agents classified as Critically Important Antimicrobials for Human Medicine, as well as antibiotics considered to pose an elevated risk for selecting resistant bacterial strains. Despite the recommendation, Snodgrass et al. (2023) reported that only 51.7% of patients receiving VAN truly had indications for its use based on the suspected diagnosis, highlighting substantial

overuse. Environmental overuse of VAN and the resulting excessive environmental exposure, often alongside third-generation cephalosporins, vancomycin-resistant enterococci (VRE) have emerged (Siegel, 2012). For this reason, VAN has been detected in the wastewater outflows from urban hospitals with a very high frequency of its occurrence (Rossmann et al., 2014; Zuccato et al., 2010). Conventional wastewater treatment plants are unable to completely remove this type of micropollutant (Deng et al., 2026). Given these concerns, the development of efficient wastewater treatment methods for the removal of such micropollutants has become a pressing priority (Szymańska et al., 2019).

Photocatalytic degradation is recognized as an effective method to remove persistent organic micropollutants. The photocatalysis begins when a semiconductor absorbs light with energy exceeding its band gap, promoting electrons (e^-) to the conduction band (CB) and generating holes (h^+) in the valence band (VB). These charge carriers (e^- , h^+) initiate oxidation and redox reactions that produce reactive oxygen species (ROS). ROS are capable of degradation of the micropollutants from the aquatic matrix (Figure 1) (Puga et al., 2024). Among various photocatalysts, titanium dioxide (TiO_2) pays special attention due to its excellent chemical stability, efficient generation of electron hole (e^-/h^+) pairs, and a band gap energy between 3.20 and 3.55 eV (Qi et al., 2025). TiO_2 -based photocatalysis has demonstrated high efficiency in degrading contaminants of emerging concern (CECs) in aquatic matrices, owing to their strong oxidative capacity, chemical stability, and effective generation of reactive oxygen species (Bayan et al., 2021; Silerio-Vázquez et al., 2024). Researchers are also increasingly investigating zinc oxide (ZnO), which offers comparable photocatalytic efficiency. Moreover, comparing the above-mentioned photocatalysts ZnO is more cost-effective, exhibits greater biocompatibility, and remains stable in environmental conditions (Bujaldón et al., 2024; Kowalska et al., 2022).

To enhance photocatalytic efficiency, the process can be coupled with sulfate radical-based advanced oxidation processes (SR-AOPs), which generate highly reactive sulfate radicals ($SO_4^{\cdot-}$) in the reaction medium. SR-AOPs have been attracting increasing attention because of their strong oxidation ability and mineralization efficiency (Long et al., 2022). Compared with HO^{\cdot} , $SO_4^{\cdot-}$ exhibit higher selectivity, extended stability, and

equal or even higher redox potential under neutral pH conditions (Li et al., 2021; Milh et al., 2021). Peroxymonosulfate (HSO_5^- , PMS) is an unsymmetrical oxidant that can partially oxidize some organic compounds. However, its direct reaction toward most pollutants is relatively low. Therefore, activation of PMS to generate $SO_4^{\cdot-}$ is necessary to achieve effective degradation (Li et al., 2021). PMS has been found as the sulfate radical precursor owing to its strong oxidative potential (2.5–3.1), making it an environmentally friendly reagent for solar-driven degradation processes. A wide range of reactive species generated during, e.g. $SO_4^{\cdot-}$, HO^{\cdot} , $\cdot O_2^-$, and 1O_2 photochemical activation degrade organic micropollutants to by-products, CO_2 , H_2O , and other mineral products as well (Jiang et al., 2022; Tian et al., 2023).

Although reaction kinetic depends on solution pH, it is also crucial to evaluate the photodegradation of VAN under various water treatment conditions, including the presence of various ions naturally occurring in aqueous matrices or extensively released through human activities. Therefore, this study aimed to investigate the impact of common anions such as nitrate (NO_3^-), chloride (Cl^-), bicarbonate (HCO_3^-), and sulfate (SO_4^{2-}), on the VAN degradation. They might inhibit photocatalysis by scavenging reactive species or interfering with the formation and subsequent reactions of radicals (e.g., hydroxyl radical (HO^{\cdot}), and sulfate radical ($SO_4^{\cdot-}$)) (He and O'Shea, 2020). Conversely, the presence of sulfate anions may enhance the degradation efficiency in persulfate-based systems. However, the influence of SO_4^{2-} on photocatalytic degradation strongly depends on the sulfate concentration and the concentration of the sulfate radical precursor as well (Honarmandrad et al., 2023).

Despite the increasing concern regarding the presence of VAN in aquatic environments, studies addressing its degradation using advanced oxidation processes (AOPs) are limited. To the best of authors' knowledge, systematic investigations on VAN removal via conventional AOPs or SR-AOPs have not yet been reported. The application of solar-driven photolysis and photocatalysis, both in the absence and presence of PMS, remains unexplored. Therefore, this study aims to fill this research gap by evaluating the efficiency and mechanisms of VAN degradation in solar-driven system using PMS as a sulfate anion radicals' precursor.

This study focused on evaluating VAN removal efficiency by solar-driven photolysis and

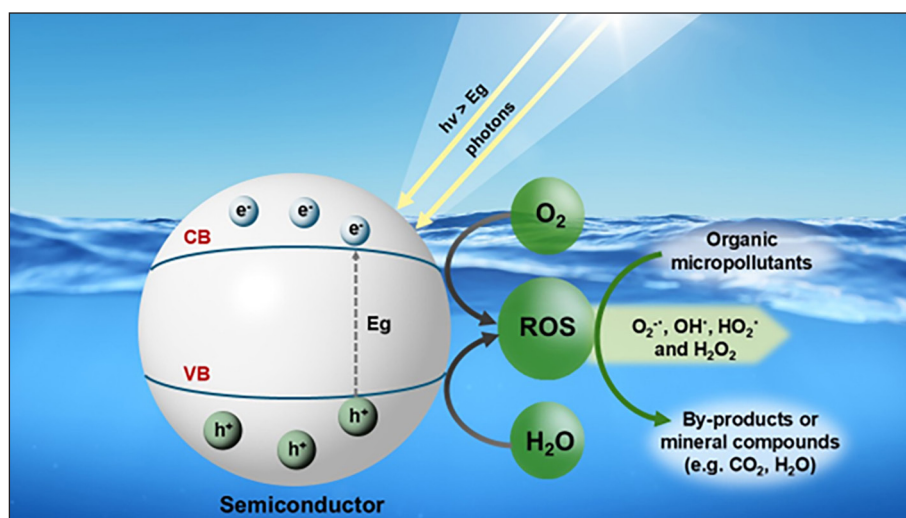


Figure 1. The schematic diagram of the photocatalytic mechanism (based on Qi et al., 2025)

photocatalysis in the presence of various anions. TiO₂-P25 and ZnO were used as photocatalysts, where TiO₂-P25 was used as a reference photocatalysts. Experiments were performed without PMS and with PMS at two concentrations (20 mg L⁻¹ and 200 mg L⁻¹), as well as with and without an additional sulfate source to assess the influence of SO₄²⁻ on the VAN degradation efficiency in PMS-assisted systems.

MATERIALS AND METHODS

The present methodology builds on and extends the approach applied in our previous works (Sowik et al., 2025a, Sowik et al., 2025b).

Reagents and chemicals

The analytical standard of VAN-HC was purchased from Sigma Aldrich (China). Peroxymonosulfate (PMS), supplied as Oxone® salt (KHSO₅·0.5KHCO₃·0.5K₂SO₄), was purchased from Sigma Aldrich (USA). Acetonitrile (ACN; HPLC grade, 99.9%) and formic acid (FA, ≥99.9%) for HPLC were obtained from VWR Chemicals (Poland). The photocatalysts: TiO₂-P25 (Titandioxid P25, 99.9%) and ZnO (99.9%), were purchased from Degussa (Germany) and Sigma Aldrich (China), respectively. TiO₂-P25 refers to the commercial Degussa (Evonik) P25 titanium dioxide – photocatalysts composed of a mixed-phase structure of anatase (~80%) and rutile (~20%). As a source of selected anions: sodium chloride (>99%) (Warchem, Poland),

sodium bicarbonate (98.8%), sodium nitrate (>99%) (Avantor, Poland), and sodium sulfate (99.9%) (Standard, Poland) were used. The mobile phase solution was prepared using ultrapure Milli-Q water (18.2 MΩ.cm) produced by a Synergy® Water Purification System (Merck, Germany).

A stock solution of VAN (10 mg L⁻¹) was prepared by dissolving 10 mg of the analytical standard in 1000 mL of Milli-Q water. Working solutions with a concentration of 2 mg L⁻¹ were freshly prepared prior to each experiment by diluting 1:5 of the stock solution also in Milli-Q water. The stock solutions were stored at 2 °C in a refrigerator and remained stable for at least three weeks.

Quantification of VAN

Vancomycin (VAN) was quantified using high-performance liquid chromatography (HPLC) with a UV variable wavelength detector (UltiMate 3000 system; Dionex Corporation, Sunnyvale, CA, USA). Separation was achieved on a C18 Hypersil™ Gold column (250 × 4.6 mm; 5 μm; Thermo Scientific, Polygen, Poland). The mobile phase consisted of 0.1% FA solution and ACN in a 77:23 (v/v) ratio. Detection was carried out at 323 nm and 280 nm, and VAN was identified at a retention time of 5.5 ± 0.8 min. The limit of quantification (LOQ) was 0.1 mg L⁻¹, corresponding to the lowest point of the calibration curve (linear regression, R² > 0.98). Data acquisition and processing were performed using Dionex Chromeleon™ 6.8 software.

Experimental set-up

Control experiments under dark conditions were conducted to evaluate the removal of VAN due to sorption onto surface of the photocatalysts (TiO₂-P25 and ZnO). The experiments were performed in glass beakers tightly covered with aluminum foil to prevent any unintended light exposure. Each experiment was carried out with a total solution volume of 250 mL and a photocatalysts concentration of 20 mg L⁻¹. Solutions containing VAN and respective photocatalysts were continuously mixed using magnetic stirrers to ensure homogeneity. The sorption experiments were conducted for 60 minutes, with samples collected at the same time intervals as those used in the photodegradation experiments.

Target laboratory-scale experiments were carried out to investigate the removal of VAN and to evaluate the influence of selected anions on VAN degradation efficiency under simulated solar irradiation. A Solarbox 1500e solar simulator (Co.fo.me. gra, Italy), equipped with a 1500 W polychromatic xenon lamp, was used to imitate the solar radiation (Figure 2). Photolytic and photocatalytic degradation of VAN were evaluated under irradiation intensity set at 500 W m⁻² (400–700 nm).

Photocatalytic experiments were performed using TiO₂-P25 and ZnO at a concentration of 20 mg L⁻¹ of each. The concentrations of the selected

anions were established based on the permissible limits for drinking water quality in Poland (Dz. U. 2017.2294), including 250 mg L⁻¹ for Cl⁻, SO₄²⁻ and HCO₃⁻ and 50 mg L⁻¹ for NO₃⁻. To assess the influence of the sulfate anion precursor, PMS was used at concentrations of 20 mg L⁻¹ and 200 mg L⁻¹. The PMS concentrations were selected based on reported ranges in literature, which span from very low to higher concentrations than those used in the present study. These values were chosen to evaluate the effect of PMS dosage on VAN degradation while recognizing further optimization is required to identify the optimal concentration for the process. Tests were also performed in the presence of sulfate ions (SO₄²⁻) at concentrations of 250 mg L⁻¹, as one of the main components of the aquatic environmental matrices. This allowed to assess the effect of PMS dosage on VAN removal and its interaction with sulfate ions.

The working volume solution was 250 mL for each experiment. The initial concentration of VAN was 2 mg L⁻¹. Although micropollutants are typically present in environmental water at ng L⁻¹ – ug L⁻¹ levels, locally elevated concentrations in the mg L⁻¹ range may occur in specific scenarios, such as pharmaceutical effluents or hospital wastewater. Moreover, the use of higher initial concentration was also necessitated by analytical detection limitations of the available equipment.



Figure 2. Solar simulator (Solarbox 1500e)

The reaction solution was placed in a glass beaker positioned inside the Solarbox and continuously stirred using a magnetic stirrer to ensure homogeneity throughout the experiment. The Solarbox is equipped with an internal mirror system that enables multiple reflections of the incident light, allowing uniform irradiation of the sample from all directions. Samples were collected at predetermined time intervals (0, 5, 10, 20, 30, 45, 60, and 90 minutes) which was the same for both photolysis and photocatalysis. All experiments were performed in duplicate.

Data analysis

The removal efficiency of VAN (R_{VAN}) was determined using to Equation 1:

$$R_{VAN} = \left(\frac{C_0 - C_t}{C_0} \right) \cdot 100\% \quad (1)$$

where: C_0 and C_t represent the initial and residual antibiotic concentrations (mg L^{-1}) at time t , respectively.

The degradation kinetics of VAN followed a pseudo first-order kinetics model, as described by Equations 2–3 (Sowik et al., 2025a):

$$\ln \left(\frac{C_t}{C_0} \right) = \exp^{-k_t t} \quad (2)$$

$$t_{1/2} = \frac{\ln 2}{k_t} \quad (3)$$

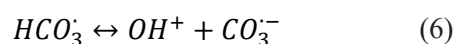
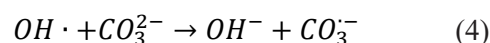
RESULTS AND DISCUSSION

Photodegradation of VAN without PMS

These experiments were carried out to determine the effect of matrix complexity on the photodegradation of VAN. Selected anions, namely nitrate (NO_3^-), chloride (Cl^-), bicarbonate (HCO_3^-), and sulfate (SO_4^{2-}), were used as chemical scavengers. Comparing the degradation efficiency of VAN in the presence and absence of each scavenger, there is a possibility to evaluate the effect of each anion on the photolytic and photocatalytic degradation of VAN.

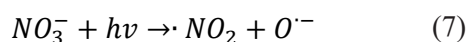
Sorption experiments under dark conditions indicated partial removal of VAN, with approximately 22% of the initial concentration adsorbed onto the ZnO surface within 60 minutes. In comparison, TiO_2 exhibited slightly higher sorption

capacity, achieving about 26% VAN removal within the same time. Photolytic degradation of VAN was inefficient, reaching only 18% removal after 90 minutes of irradiation ($t_{1/2} = 301 \text{ min}$, $k = 0.002 \text{ min}^{-1}$, $R^2 = 0.88$) (Figure 3a). The relatively low efficiency of direct photolysis can be attributed to its high molecular weight and complex polycyclic structure which limits its absorption of solar radiation and reduces the probability of direct bond cleavage. The presence of multiple aromatic rings and functional groups may promote steric hindrance, thereby decreasing the reactivity toward photogenerated species. The addition of anions slightly enhanced the degradation efficiency in the following orders: $\text{HCO}_3^- > \text{NO}_3^- > \text{Cl}^- > \text{SO}_4^{2-}$. The presence of bicarbonates increased the removal efficiency by 48 pp ($t_{1/2} = 49 \text{ min}$, $k = 0.014 \text{ min}^{-1}$, $R^2 = 0.99$). The efficiency increase in the presence of carbonates may be result of the formation of carbonate radicals ($\text{CO}_3^{\cdot-}$) by reaction with $\text{HO}\cdot$ (Equations 4–6). $\text{CO}_3^{\cdot-}$ has been found as powerful oxidant in the degradation of organic pollutants. It has high reactivity and selectivity towards pollutants with electron-rich centers and possesses a lower redox potential than $\text{HO}\cdot$. According to above, carbonates show a dual effect depending on the type of micropollutants (Rayaroth et al., 2023; Wang et al., 2021). These anions act as pH buffers, typically shifting the solution toward alkaline conditions, where many micropollutants, including antibiotics, exist in deprotonated forms that are more susceptible to photochemical transformation (Li et al., 2018). At the same time, at low concentrations, carbonates and bicarbonates act primarily as radical scavengers, converting highly reactive $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ into less reactive carbonate radicals ($\text{CO}_3^{\cdot-}$), which may reduce degradation efficiency (Ao and Liu, 2017; Liu et al., 2013). Presence of carbonates may reduce non-selective hydroxyl-based oxidation, $\text{CO}_3^{\cdot-}$ can still contribute to the degradation of specific compounds, especially under conditions where dissolved oxygen is limited (Li et al., 2018).



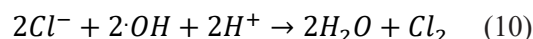
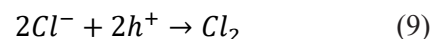
Lin et al. (2023) showed that Cl^- , SO_4^{2-} , NO_3^- had inhibitory effects on the photodegradation of

sulfamethoxazole and enrofloxacin. On the other hand, Bian and Zhang (2016) indicate that NO_3^- can be photolyzed to nitroso and negative oxygen ions and later further combined with H^+ to form $\text{HO}\cdot$ (Equations 7 and 8). The addition of the nitrate to the sulfadiazine solution resulted in enhanced degradation rate from 0.071 to 0.10. Nitrate ions, commonly present in natural waters, can act as effective photosensitizers in photocatalytic and photolytic processes. Upon UV irradiation, nitrates undergo photolysis, leading to the formation of highly reactive $\text{HO}\cdot$ through reactions involving excited nitrate species and water molecules. As a result, the degradation rate of various micropollutants often increases with increasing nitrate concentrations, reflecting enhanced $\text{HO}\cdot$ generation. However, at elevated nitrate levels, light absorption by NO_3^- may partially compete with target contaminants for photons, introducing a light-screening effect that can limit further efficiency gains (Li et al., 2018).



The highest removal efficiency of VAN among all studied processes was achieved by photocatalysis with TiO_2 , where complete VAN removal was observed after only 20 minutes ($t_{1/2} = 2.93$ min, $k = 0.237$ min⁻¹, $R^2 = 0.97$) (Figure 3b). A comparable result was obtained in the presence of NO_3^- , indicating that nitrate slightly affect the process ($t_{1/2} = 5.04$ min, $k = 0.138$ min⁻¹, $R^2 = 0.95$). In contrast, other anions (Cl^- , HCO_3^- , SO_4^{2-}) reduced the photodegradation rate, and complete removal was observed after 45 minutes. Chlorides ions can act as effective scavengers of photogenerated charge carriers and reactive oxygen species, including valence band holes (h^+), $\text{HO}\cdot$, and $\cdot\text{O}_2^-$, leading to the formation of reactive chlorine species (RCS) such as Cl_2 or $\text{HOCl}\cdot$ (Equations 9–11). While these reactions often reduce the availability of highly reactive $\text{HO}\cdot$ and slow overall mineralization, the generated RCS may selectively oxidize electron-rich organic micropollutants, what can result in faster removal of certain parent compounds. However, chloride ions may also compete with pollutants for active sites on the photocatalysts surface, further limiting degradation efficiency (Sun et al., 2024; Chi et al., 2020). For example, presence of Cl^- at certain value enhanced removal of dye pollutants. But decreasing the overall

photodegradation efficiency occurred when the Cl^- concentration was below the appropriate value (Wang et al., 2011). Moreover, at low pH values, chloride ions can form inner-sphere surface complexes with TiO_2 . According to that, Cl^- competes with hydroxyl ions and water molecules for active sites on the TiO_2 surface, leading to reduced $\text{HO}\cdot$ radical generation (Puga et al., 2024; Villegas-Guzman et al., 2017).



ZnO-based photocatalysis showed a significant decrease in VAN removal efficiency compared to TiO_2 -based processes ($t_{1/2} = 8.57$ min, $k = 0.081$ min⁻¹, $R^2 = 0.86$) (Figure 3c). Complete removal was achieved after 90 minutes. The presence of SO_4^{2-} further decreased efficiency, resulting in only 89% removal of VAN ($t_{1/2} = 27$ min, $k = 0.025$ min⁻¹, $R^2 = 0.97$). Conversely, the presence of the other anions (Cl^- , HCO_3^- and NO_3^-) positively influenced the photocatalytic degradation of VAN, where complete removal was obtained after 45 minutes for Cl^- ($t_{1/2} = 9.16$ min, $k = 0.076$ min⁻¹, $R^2 = 0.86$) and HCO_3^- ($t_{1/2} = 11.99$ min, $k = 0.058$ min⁻¹, $R^2 = 0.93$) and after 60 minutes for NO_3^- ($t_{1/2} = 11.46$ min, $k = 0.0605$ min⁻¹, $R^2 = 0.94$). The presence of NO_3^- in photocatalysis with ZnO process slightly improved VAN degradation, which may be attributed to nitrate photolysis under irradiation, generating $\cdot\text{NO}_2$ and O^- radicals that can further react to form $\cdot\text{OH}$ (Bian and Zhang, 2016). However, previous studies have reported other effects of nitrate, with Lin et al. (2023) showing inhibitory behavior. Therefore, the observed enhancement should be considered as a process-specific effect and interpreted with caution. Lower efficiency of ZnO-based photocatalysis in the presence of SO_4^{2-} might be result of too high concentration of SO_4^{2-} beyond the optimum value. Under such conditions, SO_4^{2-} can interact with the ZnO surface, contributing to catalyst site deactivation and as result reduce the degradation efficiency of organic contaminants (Mirzaei et al., 2016). In turn, increased efficiency in other cases may be the result of the formation of new reactive radicals.

The presence of Cl^- and other ions, such as SO_4^{2-} , NO_3^- and HCO_3^- are commonly studied as interfering species, where their presence together with the micropollutant, can influence the efficiency

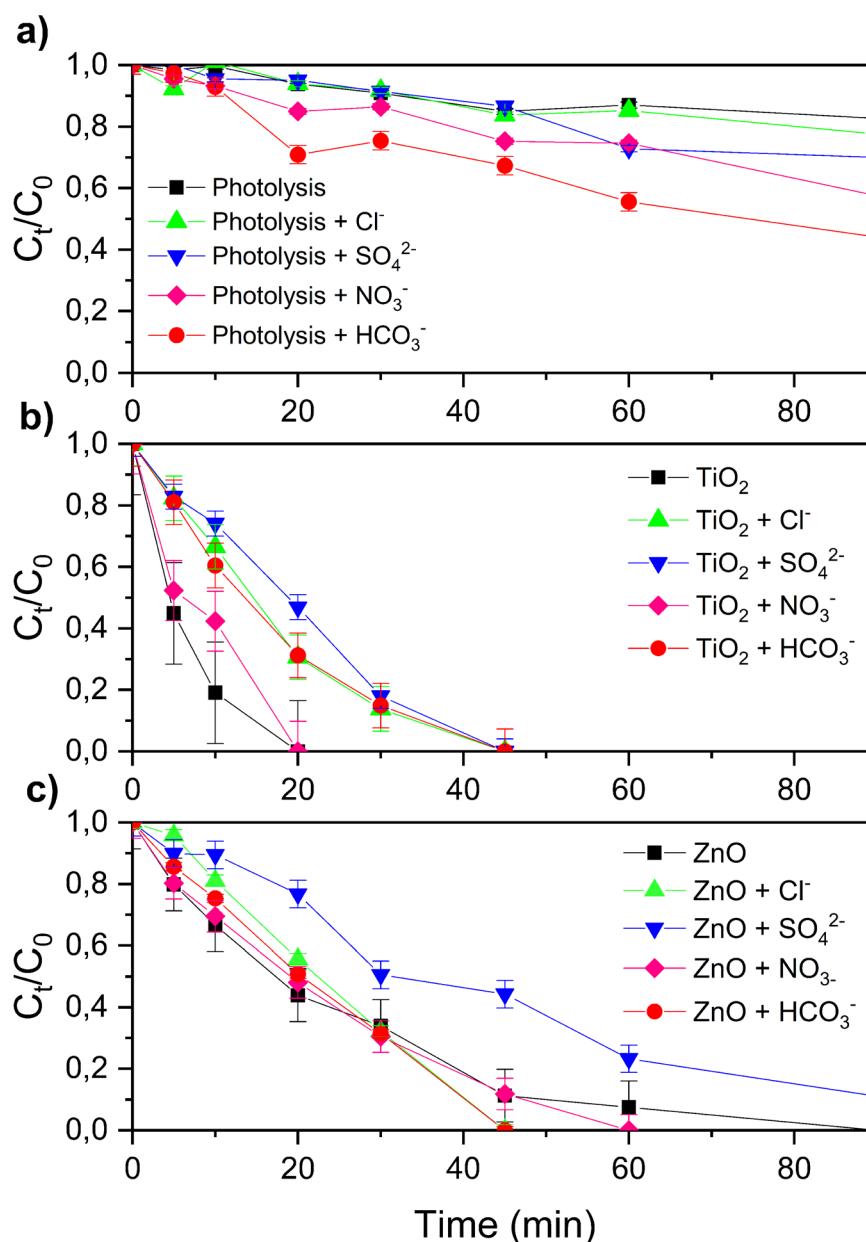


Figure 3. Photodegradation of VAN a) photolysis; b) TiO_2 -based photocatalysis; c) ZnO -based photocatalysis without and in the presence of studied anions ($Cl^- = 250 \text{ mg L}^{-1}$; $SO_4^{2-} = 250 \text{ mg L}^{-1}$; $NO_3^- = 50 \text{ mg L}^{-1}$ and $HCO_3^- = 250 \text{ mg L}^{-1}$)

of photocatalytic degradation. Confirming photocatalytic activity in the presence of these ions is critical for the development of applications in wastewater treatment systems (Hong et al., 2022).

The error bars represent standard deviations calculated from duplicate experiments and are provided to indicate experimental reproducibility.

Photodegradation of VAN with PMS

Anions might affect the efficiency of antibiotic photodegradation using PMS in aquatic matrices in

different ways. Understanding their impact is crucial, as these anions naturally occurred in aquatic environments can either inhibit or enhance micropollutants removal. The overall effect depends on several factors, including the type of process, the concentration of the target micropollutant, the type and dosage of oxidant, and concentration of present anions. Many studies have examined the effects of Cl^- , NO_3^- , $H_2PO_4^-$, and HCO_3^- on photodegradation of CECs (He and O’Shea, 2020; Nie et al., 2019; Raczak et al., 2025; Yang et al., 2018).

However, there is still incomplete knowledge of the influence of SO_4^{2-} on PMS-based photodegradation processes that rely on the generation of sulfate radical anions ($\text{SO}_4^{\cdot-}$).

The presence of $\text{SO}_4^{\cdot-}$ negatively affects the photodegradation of VAN. In all cases, the removal efficiency decreased in comparison to PMS alone, with the lowest degradation rate observed for the photolysis. Complete VAN removal was achieved after 60 and 20 minutes for PMS concentrations of 20 mg L⁻¹ (PMS20) ($t = 14.75$ min, $k = 0.047$ min⁻¹, $R^2 = 0.96$) and 200 mg L⁻¹ (PMS200) ($t = 4.74$ min, $k = 0.146$ min⁻¹, $R^2 = 0.97$), respectively (Figure 4a). Addition of sulfates results in lower degradation efficiency, where complete degradation was obtained after 45 minutes with PMS200 ($t = 9.43$ min, $k = 0.074$ min⁻¹, $R^2 = 0.93$), while approximately 73% of removal was achieved within 90 minutes at the lower PMS concentration ($t = 52.12$ min, $k = 0.013$ min⁻¹, $R^2 = 0.96$). The smallest differences between processes were observed for the TiO₂-based photocatalysis, where the presence of $\text{SO}_4^{\cdot-}$ caused only a minor decrease in efficiency, especially at the higher PMS concentration (Figure 4b). In contrast, the greatest differences occurred in the ZnO-based processes, where $\text{SO}_4^{\cdot-}$ reduced degradation efficiency at both PMS levels, indicating a stronger inhibitory effect on ZnO than on TiO₂. Moreover, increasing the PMS concentration significantly enhanced VAN removal during all processes. However, the highest improvement was for the photolytic and ZnO-based processes, which showed much efficient degradation at PMS200 compared to PMS20. In contrast, photocatalysis with TiO₂, already highly efficient at the lower PMS concentration, exhibited the smallest relative enhancement upon increasing PMS dosage.

Error bars represent standard deviations calculated from duplicate experiments and are provided to indicate experimental reproducibility. The highest efficiency of VAN removal was achieved by photocatalysis using TiO₂ and PMS at a concentration of 200 mg L⁻¹ ($t_{1/2} = 2.62$ min, $k = 0.264$ min⁻¹, $R^2 = 0.93$). Decrease of removal efficiency was observed only at the lower PMS concentration in the presence of sulfates ($t_{1/2} = 6.94$ min, $k = 0.099$ min⁻¹, $R^2 = 0.98$). In turn, in any case of the ZnO-based process, the presence of SO_4^{2-} reduces the efficiency of the VAN photodegradation process. The observed effects in ZnO/PMS processes are likely related to PMS activation, although the contribution of Zn^{2+}

release cannot be excluded and was not directly investigated (Kong et al., 2021). In the process with a lower PMS concentration (20 mg L⁻¹), complete VAN removal occurred after 30 and 45 minutes, respectively. However, in the processes with PMS at a concentration of 200 mg L⁻¹, complete removal was achieved within 20 and 30 minutes, both for the process without and in the presence of SO_4^{2-} .

The enhanced degradation efficiency observed in PMS-assisted photocatalytic processes may suggest that VAN removal is predominantly ruled by non-selective radical reactions, involving $\text{SO}_4^{\cdot-}$, which are more effective in attacking structurally complex antibiotic molecules. Moreover, pH is one of the most influential parameters in PMS-based processes, as it rules PMS stability, speciation, and the relative contribution of reactive radicals, directly affecting antibiotics degradation efficiency. Under acidic conditions (pH < 7) – like in our case, where pH of most of the processes is in the range 3.4–6.1, $\text{SO}_4^{\cdot-}$ are the dominant reactive species. However, at very low pH values, excess H⁺ may act as radical scavengers, potentially reducing degradation rates. In neutral and mildly alkaline conditions can be converted into HO• may coexist, whereas at higher pH values (pH > 9), sulfate radicals can be converted into hydroxyl radicals through reactions with OH⁻, often resulting in lower oxidation efficiency due to the weaker oxidizing potential and shorter lifetime of HO•. Additionally, alkaline conditions may promote PMS decomposition pathways leading to the formation of H₂O₂, O₂^{•-}, ¹O₂, which can further modify degradation mechanisms. pH also affects catalysts behavior, including metal speciation, surface hydroxylation, and the formation of secondary radical scavengers such as bicarbonate and carbonate ions. Therefore, although PMS-based processes can operate over a wide pH range, the balance between $\text{SO}_4^{\cdot-}$ and HO• and degradation efficiency are strongly process-dependent, highlighting the need for pH-specific optimization tailored to the target antibiotic (Ghanbari et al., 2016; Honarmandrad et al., 2023 Wang et al., 2018).

Sulfate ions exhibit a limited influence on degradation efficiency in PMS-based processes. Previous studies on the degradation of pharmaceuticals such as sulfamethoxazole and atenolol have shown that SO_4^{2-} neither significantly inhibits nor promotes pollutant removal. This behavior may be attributed to the relatively low reactivity of sulfate ions toward both sulfate and hydroxyl radicals

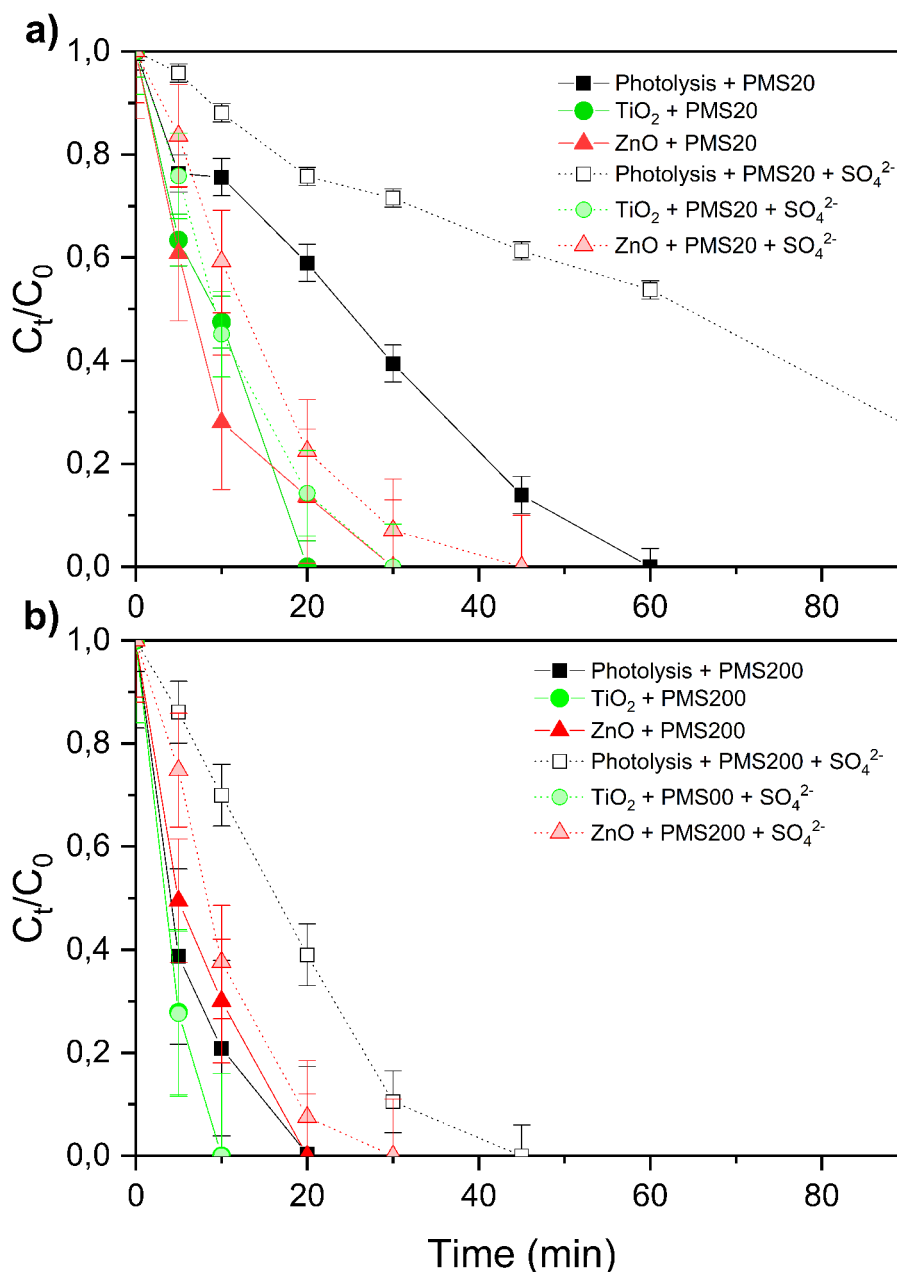


Figure 4. Photodegradation of VAN in the absence and presence of SO_4^{2-} with PMS concentration at a) 20 mg L^{-1} and b) 200 mg L^{-1}

compared to other common inorganic anions. Consequently, SO_4^{2-} is not considered an effective radical scavenger under typical UV/PMS conditions, and its presence does not substantially deteriorate process performance. Only in specific systems and at low concentrations may SO_4^{2-} slightly enhance degradation, whereas higher concentrations tend to exhibit a neutral or mildly inhibitory effect (Ao and Liu, 2017; Liu et al., 2013). From the other hand, Honarmandrad et al. (2023) reported that sulfate ions negatively affect $\text{SO}_4^{\cdot-}$ and $\text{HO}\cdot$, because SO_4^{2-} can suppress those radicals and may

also chelate metal oxide. Moreover, the scavenging effect of SO_4^{2-} might become stronger with increasing SO_4^{2-} concentration, which in turn led to the reduction of micropollutants degradation rate at high SO_4^{2-} concentration. Similar results were observed during the degradation of cefixime using a UV/PMS system, where low concentrations of SO_4^{2-} ions decreased the photodegradation efficiency (Khazaei et al., 2019). Nevertheless, our recent work demonstrated that the PMS/Vis system achieved much higher removal of VAN in a more complex matrix, such as real municipal wastewater

effluent (MWW). It is highlighting that, despite the potential inhibitory role of SO_4^{2-} , processes using PMS as oxidant can perform very effectively under realistic environmental conditions (Sowik et al., 2025). Although direct reactions between anions and PMS were not assessed in this study, we recognize their potential significance for PMS activation and radical generation.

CONCLUSIONS

VAN photodegradation in the solar light spectrum is influenced by several factors such as the type of process, photocatalysts and its concentration, and the dosage of oxidant. Increasing PMS concentration enhanced the photodegradation rate of VAN, due to higher generation of more reactive oxygen species. However, the presence of SO_4^{2-} negatively affected the photodegradation of VAN, leading to a significant depletion of removal efficiency. In case of photocatalysts examined, TiO_2 outperforms ZnO , showing higher photodegradation efficiency under artificial sunlight. Moreover, the study confirms that anions naturally present in aquatic environments can inhibit or promote photodegradation, based on their concentration and chemical characteristics. These effects do not follow a simple trend. Anions interact with oxidant like PMS and compete for reactive sites on the catalysts, potentially altering the entire degradation mechanism. In real environmental matrices, the coexistence of various ions and organic matter introduces additional competitive and synergistic reactions, leading to results that may differ from those obtained with single-ion model systems. According to the complexity of relations between reactions, it is necessary to evaluate advanced oxidation processes in the environmental samples, such as real wastewater, to properly assess and understand the possibility of implementation of photocatalytic technologies for micropollutants removal in wastewater treatment plants. Overall, the results indicate that the proposed PMS-based process operating under visible-light irradiation has strong potential for application in wastewater treatment plants due to its high efficiency and low energy demand, as it does not require UV light. Nevertheless, further studies should include analysis of transformation products, toxicity assessment of post-process solutions, and a systematic evaluation of pH and anion synergy. Those are necessary to fully assess practical applicability of the technology.

Acknowledgments

Funding was provided by the Silesian University of Technology under the young researchers' grant BKM-601/RIE7/2025. Katarzyna Kowalska also acknowledges financial support from the Polish Ministry of Science and Higher Education (grant no. 08/070/BK_25/0038) and 08/070/BK_26/0044 (BK-241/RIE7/2026).

REFERENCES

1. Ao, X., Liu, W. Degradation of sulfamethoxazole by medium pressure UV and oxidants: Peroxymonosulfate, persulfate, and hydrogen peroxide, *Chemical Engineering Journal* 2017; 313, 629–637. <http://dx.doi.org/10.1016/j.cej.2016.12.089>
2. Bayan, E. M., Pustovaya, L. E., Volkova, M. G. Recent advances in TiO_2 -based materials for photocatalytic degradation of antibiotics in aqueous systems. In *Environmental Technology and Innovation* 2021; 24. Elsevier B.V. <https://doi.org/10.1016/j.eti.2021.101822>
3. Bian, X., Zhang, J. Photodegradation of sulfadiazine in aqueous solution and the affecting factors. *Journal of Chemistry*, 2016. <https://doi.org/10.1155/2016/8358960>
4. Bujaldón, R., Benamara, M., Dhahri, R., Gómez, E., Serrà, A. Attuning doped ZnO-based composites for an effective light-driven mineralization of pharmaceuticals via PMS activation. *Chemosphere*, 2024; 357. <https://doi.org/10.1016/j.chemosphere.2024.142127>
5. Chi, Y., Xu, S., Li, M., He, M., Yu, H., Li, L., Yue, Q., Gao, B. Effective blockage of chloride ion quenching and chlorinated by-product generation in photocatalytic wastewater treatment, *Journal of Hazardous Materials*, 2020; 326; 122670. <https://doi.org/10.1016/j.jhazmat.2020.122670>
6. Deng, K., Zhao, L., Li, M., Liang, X., Wang, X., Han, J., Zhang, Y., Kiyama, R., Dong, S. Antibiotics in northern China's largest shallow wetland lake: Occurrence, bioaccumulation, ecological risk, and indicator screening. *Journal of Environmental Sciences (China)*, 2026; 159, 233–241. <https://doi.org/10.1016/j.jes.2025.04.012>
7. Dz.U. 2017 poz. 2294 Rozporządzenie Ministra Zdrowia z dnia 7 grudnia 2017 r. w sprawie jakości wody przeznaczonej do spożycia przez ludzi.
8. WHO Regional Office for Europe and European Centre for Disease Prevention and Control. Surveillance of antimicrobial resistance in Europe, 2024 data: executive summary. Copenhagen: WHO Regional Office for Europe; 2025.
9. Felis, E., Kalka, J., Sochacki, A., Kowalska, K.,

- Bajkacz, S., Harnisz, M., Korzeniewska, E. Antimicrobial pharmaceuticals in the aquatic environment - occurrence and environmental implications. *European Journal of Pharmacology*, August, 2019; 172813. <https://doi.org/10.1016/j.ejphar.2019.172813>
10. Ghanbari, F., Moradi, Mahsa. Application of peroxymonosulfate and its activation methods for degradation of environmental organic pollutants: Review. *Chemical Engineering Journal*, 2016; 310: 1, 41–62. <https://doi.org/10.1016/j.cej.2016.10.064>
 11. He, X., O’Shea, K. E. Selective oxidation of H1-antihistamines by unactivated peroxymonosulfate (PMS): Influence of inorganic anions and organic compounds. *Water Research*, 2020. 186. <https://doi.org/10.1016/j.watres.2020.116401>
 12. Honarmandrad, Z., Sun, X., Wang, Z., Naushad, M., Boczkaj, G. Activated persulfate and peroxymonosulfate based advanced oxidation processes (AOPs) for antibiotics degradation - A review. *Water Resources and Industry*, 2023; 29. <https://doi.org/10.1016/j.wri.2022.100194>
 13. Hong, J., Cho, Ki., Presser, V., Su, X. Recent advances in wastewater treatment using semiconductor photocatalysts. *Green and Sustainable Chemistry*. 2022; 36, 100644. <https://doi.org/10.1016/j.cogsc.2022.100644>
 14. Jiang, Z. R., Li, Y., Zhou, Y. X., Liu, X., Wang, C., Lan, Y., Li, Y. Co₃O₄-MnO₂ nanoparticles moored on biochar as a catalyst for activation of peroxymonosulfate to efficiently degrade sulfonamide antibiotics. *Separation and Purification Technology*, 2022; 281(June 2021), 119935. <https://doi.org/10.1016/j.seppur.2021.119935>
 15. Khazaei, R., Rahmani, A., Seidmohammadi, A., Faradmal, J., Leili, M. Evaluation of the efficiency of photocatalytic uv/ peroxymonosulfate process in the removal of cefexime antibiotic from aqueous solutions. *Scientific Journal of Kurdistan University of Medical Sciences*, 2019; 24(4), 22–40. <https://doi.org/10.29252/sjku.24.4.22>
 16. Kong, L., Fang, G., Fang, Z., Zou, Y., Zhu, F., Zhou, D., Zhan, J. Peroxymonosulfate activation by localized electrons of ZnO oxygen vacancies for contaminant degradation. *Chemical Engineering Journal*, 2021; 416, 128996. <https://doi.org/10.1016/j.cej.2021.128996>
 17. Kowalska, K., Roccamante, M., Cabarera Reina, A., Plaza-Bolanos, P., Oller, I., Mallato, S. Pilot-scale removal of microcontaminants by solar-driven photo-Fenton in treated municipal effluents: selection of operating variables based on lab-scale experiments. *Journal of Environmental Chemical Engineering*, 2022; 9(1), 104788. <https://doi.org/10.1016/j.jece.2020.104788>
 18. Li, C., Zhang, D., Peng, J., Li, X.. The effect of pH, nitrate, iron (III) and bicarbonate on photodegradation of oxytetracycline in aqueous solution. *Journal of Photochemistry and Photobiology A: Chemistry*, 2018; 356, 239–247. <https://doi.org/10.1016/j.jphotochem.2018.01.004>
 19. Li, D., Zhang, N., Yuan, R., Chen, H., Wang, F., Zhou, B. Effect of wavelengths on photocatalytic oxidation mechanism of sulfadiazine and sulfamethoxazole in the presence of TiO₂. *Journal of Environmental Chemical Engineering*, 2021; 9(5), 106243. <https://doi.org/10.1016/j.jece.2021.106243>
 20. Lin, X., Zhou, W., Li, S., Fang, H., Fu, S., Xu, J., Huang, J. Photodegradation of sulfamethoxazole and enrofloxacin under UV and simulated solar light irradiation. *Water (Switzerland)*, 2023; 15(3). <https://doi.org/10.3390/w15030517>
 21. Liu, X., Zhang, T., Zhou, Y., Fang, L., Shao, Y. Degradation of atenolol by UV/peroxymonosulfate: Kinetics, effect of operational parameters and mechanism. *Chemosphere*, 2013; 93, 2717–2724. <http://dx.doi.org/10.1016/j.chemosphere.2013.08.090>
 22. Long, L., Bai, C., Zhou, X., Zhang, S., Zhang, Y., Chen, C., He, J., Song, C., Yang, G. A novel strategy for promoting PMS activation: Enhanced utilization of side reactions. *Separation and Purification Technology*, 2022; 297(June), 121432. <https://doi.org/10.1016/j.seppur.2022.121432>
 23. Milh, H., Cabooter, D., Dewil, R. Role of process parameters in the degradation of sulfamethoxazole by heat-activated peroxymonosulfate oxidation: Radical identification and elucidation of the degradation mechanism. *Chemical Engineering Journal*, 2021; 422(March), 130457. <https://doi.org/10.1016/j.cej.2021.130457>
 24. Mirzaei, A., Chen, Z., Haghghat, F., Yerushalmi, L. Removal of pharmaceuticals and endocrine disrupting compounds from water by zinc oxide-based photocatalytic degradation: A review. *Sustainable Cities and Society*, 2016; 27, 407–418. <https://doi.org/10.1016/j.scs.2016.08.004>
 25. Nie, M., Zhang, W., Yan, C., Xu, W., Wu, L., Ye, Y., Hu, Y., Dong, W. Enhanced removal of organic contaminants in water by the combination of peroxymonosulfate and carbonate. *Science of the Total Environment*, 2019; 647, 734–743. <https://doi.org/10.1016/j.scitotenv.2018.08.065>
 26. Puga, F., Navio, J. A., Hidalgo, M. C. A critical view about use of scavengers for reactive species in heterogeneous photocatalysis. *Applied Catalysis A: General* 2024; 685. <https://doi.org/10.1016/j.apcata.2024.119879>
 27. Qi, X. N., Zhu, Y. Q., Yuan, H., Xu, Z. X., Guo, H. R., Li, Z. L. Latest progress based on doped-TiO₂ photocatalysis materials: Fabrication and applications. *Journal of Water Process Engineering* 2025; 78. <https://doi.org/10.1016/j.jwpe.2025.108791>
 28. Raczak, K. B., Silvestri, D., Myslik, P., Marton, P.,

- Salava, M., Padil, V. V. T., Černík, M., Xiao, R., Dionysiou, D. D., Waclawek, S. Sulfate radical-mediated degradation of vancomycin: Kinetics, mechanism, and toxicity evaluation. Separation and Purification Technology, 2025; 379. <https://doi.org/10.1016/j.seppur.2025.134935>
29. Rayaroth, M. P., Boczkaj, G., Aubry, O., Aravind, U. K., Aravindakumar, C. T. Advanced oxidation processes for degradation of water pollutants—ambivalent impact of carbonate species: A review. Water (Switzerland) 2023; 15(8). <https://doi.org/10.3390/w15081615>
 30. Rossmann, J., Schubert, S., Gurke, R., Oertel, R., Kirch, W. Simultaneous determination of most prescribed antibiotics in multiple urban wastewater by SPE-LC-MS/MS. Journal of Chromatography B: Analytical Technologies in the Biomedical and Life Sciences, 2014; 969, 162–170. <https://doi.org/10.1016/j.jchromb.2014.08.008>
 31. Silerio-Vázquez, F. de J., González-Burciaga, L. A., Antileo, C., Núñez-Núñez, C. M., Proal-Nájera, J. B. Photocatalytic degradation of antibiotics in water via TiO₂-x: Research needs for technological advancements. Journal of Hazardous Materials Advances 2024; 16. <https://doi.org/10.1016/j.hazadv.2024.100506>
 32. Siegel J.D., Pediatric infection prevention and control. In: Sarah S. Long (Ed.), Principles and Practice of Pediatric Infectious Diseases (Fourth Edition), 2012; 9–24.
 33. Sowik, P., Kowalska, K., Felis, E. Assessment of photodegradation efficiency of selected antibiotics from WHO watch group in environmental samples. Journal of Ecological Engineering. 2025a; 26(5), 159–169. <https://doi.org/10.12911/22998993/200062>
 34. Sowik P., Wilk J., Kowalska K., Bajkacz S., Felis E. Photocatalysis degradation of ciprofloxacin via peroxy-monosulfate activation in aquatic matrices: Comparative analysis, kinetics, and transformation pathways. Desalination and Water Treatment 2025b; 321, 101001 <https://doi.org/10.1016/j.dwt.2025.101001>
 35. Snodgrass, D., White, K.E., Azar, A., Sudhindra, P. Utilization of vancomycin in the emergency department at a community hospital and implications of its use. Open Forum Infectious Diseases 2023; 10, 2276. <https://doi.org/10.1093/ofid/ofad500.1898>
 36. Sun, Y., Sun, W., Li, Y., Dong, N., Yu, H., Yin, W., Zhu, F., Gao, B., Xu, S. Effective inhibition of chloride ion interference in photocatalytic process by negatively charged molecularly imprinted photocatalyst: Behavior and mechanism. Water Research, 2024; 262, 122040. <https://doi.org/10.1016/j.watres.2024.122040>
 37. Szymańska, U., Wiergowski, M., Soltyszewski, I., Kuzemko, J., Wiergowska, G., Woźniak, M. K. Presence of antibiotics in the aquatic environment in Europe and their analytical monitoring: Recent trends and perspectives. Microchemical Journal, 2019; 147(March), 729–740. <https://doi.org/10.1016/j.microc.2019.04.003>
 38. Tian, N., Giannakis, S., Akbarzadeh, L., Hasanvandian, F., Dehghanifard, E., Kakavandi, B. Improved catalytic performance of ZnO via coupling with CoFe₂O₄ and carbon nanotubes: A new, photocatalysis-mediated peroxy-monosulfate activation system, applied towards Cefixime degradation. Journal of Environmental Management, 2023; 329(December 2022), 117022. <https://doi.org/10.1016/j.jenvman.2022.117022>
 39. Villegas-Guzman, P., Hofer, F., Silva-Agreto, J., Torres-Palma, R. A. Role of sulfate, chloride, and nitrate anions on the degradation of fluoroquinolone antibiotics by photoelectro-Fenton. Environmental Science and Pollution Research, 2017; 24(36), 28175–28189. <https://doi.org/10.1007/s11356-017-0404-5>
 40. Wang, Z., Yuan, R., Guo, Y., Xu, L., Liu, J. Effects of chloride ions on bleaching of azo dyes by Co²⁺/oxone reagent: Kinetic analysis. Journal of Hazardous Materials, 2011; 190, 1083–1087. <https://doi.org/10.1016/j.jhazmat.2011.04.016>
 41. Wang, J., Wang J. Activation of persulfate (PS) and peroxy-monosulfate (PMS) and application for the degradation of emerging contaminants. Chemical Engineering Journal, 2018; 334, 1502–1517. <https://doi.org/10.1016/j.cej.2017.11.059>
 42. Wang, J., Wang, K., Zhang, L., Guo, Y., Guo, Z., Sun, W., Ye, Z., Niu, J. Mechanism of bicarbonate enhancing the photodegradation of β-blockers in natural waters. Water Research, 2021; 197. <https://doi.org/10.1016/j.watres.2021.117078>
 43. World Health Organization, AWaRe classification of antibiotics for evaluation and monitoring of use, 2023, WHO/MHP/HPS/EML/2023.04
 44. Yang, F., Huang, Y., Fang, C., Xue, Y., Ai, L., Liu, J., Wang, Z. Peroxy-monosulfate/base process in saline wastewater treatment: The fight between alkalinity and chloride ions. Chemosphere, 2018; 199, 84–88. <https://doi.org/10.1016/j.chemosphere.2018.02.023>
 45. Zhang, L., Du, S., Liu, D., Dong, D., Zhang, W., Guo, Z. Antibiotics in fish caught from ice-sealed waters: Spatial and species variations, tissue distribution, bioaccumulation, and human health risk. Science of the Total Environment, 2022; 821. <https://doi.org/10.1016/j.scitotenv.2022.153354>
 46. Zuccato, E., Castiglioni, S., Bagnati, R., Melis, M., Fanelli, R. Source, occurrence and fate of antibiotics in the Italian aquatic environment. Journal of Hazardous Materials, 2010; 179(1–3), 1042–1048. <https://doi.org/10.1016/j.jhazmat.2010.03.110>